

Final Draft
of the original manuscript:

Soerensen, A.L.; Jacob, D.J.; Streets, D.G.; Witt, M.L.I.; Ebinghaus, R.;
Mason, R.P.; Andersson, M.; Sunderland, E.M.:

**Multi-decadal decline of mercury in the North Atlantic
atmosphere explained by changing subsurface seawater
concentrations**

In: Geophysical Research Letters (2012) AGU

DOI: 10.1029/2012GL053736

Multi-decadal decline of mercury in the North Atlantic atmosphere explained by changing subsurface seawater concentrations

Anne L. Soerensen^{a,b}, Daniel J. Jacob^b, David Streets^c, Melanie Witt^d, Ralf Ebinghaus^e, Robert P. Mason^f, Maria Andersson^f, Elsie M. Sunderland^{a,b}*

^aHarvard School of Public Health, Department of Environmental Health, Boston MA, 02215, USA

^bHarvard University, School of Engineering and Applied Sciences, Cambridge MA, 02138, USA

^cDecision and Information Sciences Division, Argonne National Laboratory, 9700 S. Cass Avenue, Argonne, IL, 60439, USA

^dUniversity of Oxford, Department of Earth Sciences, South Park Road, Oxford, OX1 3AN, UK

^eHelmholtz-Zentrum Geesthacht, Institute of Coastal Research, Max-Planck-Strasse, 21502 Geesthacht, Germany

^fUniversity of Connecticut, Department of Marine Sciences, 1080 Sennecossett Road, Groton, CT, 0634, USA

Abstract

We analyze 1977-2010 trends in atmospheric mercury (Hg) from 21 ship cruises in the North Atlantic (NA) and 15 in the South Atlantic (SA). We find a steep 1990-2010 decline of $-0.046 \pm 0.010 \text{ ng m}^{-3} \text{ a}^{-1}$ ($-2.4\% \text{ a}^{-1}$) in NA surface air (steeper than at NH land sites) but no significant decline in the SA. Surface water Hg^0 measurements in the NA show a decline of $-5.7\% \text{ a}^{-1}$ since 1999, while the few subsurface ocean data available show $\sim 85\%$ decline from 1980 to

present. We use a coupled global atmosphere-ocean model to show that the decline in the NA atmosphere can be explained by decreasing evasion from the ocean driven by declining subsurface water Hg concentrations. We propose that past historical enrichment of the NA ocean was caused by enhanced riverine and wastewater discharges of Hg.

1. Introduction

Mercury (Hg) is emitted to the atmosphere by natural processes (crustal degassing, volcanoes) and also by human activities (fuel combustion, industry, mining). This atmospheric Hg deposits to the surface, cycles through ecosystems, may be re-emitted to the atmosphere, and is eventually incorporated into more stable reservoirs in the soil and the deep ocean.

Accumulation of mercury in ecosystems and subsequent human exposure through fish consumption is a major environmental concern [Mergler *et al.*, 2007; Mahaffey *et al.*, 2009].

Total gaseous mercury (TGM) has been measured on ship cruises since the 1970s [Temme *et al.*, 2003] and at long-term land monitoring sites since the 1990s [Temme *et al.*, 2007; Slemr *et al.*, 2011; Toerseth *et al.*, 2012]. Several recent studies have reported a declining Hg trend. Slemr *et al.* [2011] found that atmospheric Hg decreased worldwide by 20-38% over the 1996-2009 period, based on long-term data from Mace Head (Ireland) and Cape Point (South Africa), as well as five cruises over the Atlantic Ocean. Statistically significant declines for the 1995-2009 period have also been reported for Mace Head [Ebinghaus *et al.*, 2011], Alert in Arctic Canada [Cole and Steffen, 2010], and St. Anicet and Kuujuarapik in eastern Canada [Cole *et al.*, 2011]. These decreases are inconsistent with the Hg emission inventory from Streets *et al.* [2011], which finds a global increase of 30% over the past decade driven by rising Asian

emissions. Atmospheric Hg^0 is sufficiently long-lived that emission trends should be reflected in atmospheric observations at least on a hemispheric scale [Corbitt *et al.*, 2011].

The observed decline in atmospheric Hg could reflect a decrease in Hg re-emission from geochemical reservoirs to the atmosphere [Slemr *et al.*, 2011]. In particular, high Hg in subsurface seawater of the North Atlantic (NA) [Laurier *et al.*, 2004] has been attributed to a historical legacy of 20th-century anthropogenic emissions [Sunderland and Mason, 2007] and can explain the high surface air concentrations observed in NA cruises [Soerensen *et al.*, 2010b]. A decline in this subsurface loading would cause atmospheric concentrations to decrease.

Here we present an analysis of 1977-2010 Hg trends in the atmosphere and surface ocean using the full ensemble of cruise data available for the NA and South Atlantic (SA). We find a steep decline in the NA since 1990 but no significant trend in the SA. We show that the NA decline can be explained by historical enrichment of the subsurface ocean and propose that past wastewater discharges could be a major factor in this enrichment.

2. North and South Atlantic trends

We analyzed atmospheric data from 21 ship cruises in the NA (5°N-65°N), and 15 in the SA (70°S-11°N) (Figure 1) between 1977 and 2010. We know of no cruise data between 1981 and 1989. We separate NA from SA by the location of the Intertropical Convergence Zone (ITCZ) at the time of the cruise. Atmospheric Hg is measured as either TGM $\equiv \text{Hg}^0 + \text{Hg}^{\text{II}}$ or as Hg^0 . Hg^{II} accounts for only a few percent of TGM in the remote MBL [Gustin and Jaffe, 2010], thus we do not distinguish between the two in our discussion.

Figure 1 shows the 1977-2010 surface air concentrations for the ensemble of cruises, separately for the NA and the SA. Also shown are the 1996-2009 data for Mace Head and Cape Point. There are no significant trends between the late 1970s and the early 1990s for either the

NA or SA. For 1990-2009, we find a significant decrease in the NA of $-0.046 \pm 0.010 \text{ ng m}^{-3} \text{ a}^{-1}$ (\pm standard error) or $-2.4\% \text{ a}^{-1}$ ($p < 0.001$, $n = 18$), steeper than the 1996-2009 decrease reported by *Slemr et al.* [2011] and *Ebinghaus et al.* [2011] for Mace Head ($-0.024 \pm 0.005 \text{ ng m}^{-3} \text{ a}^{-1}$ and $-0.028 \pm 0.01 \text{ ng m}^{-3} \text{ a}^{-1}$, respectively). Accounting for the seasonal variation of Hg over the NA [*Soerensen et al.*, 2010a] does not change this result. We find no significant trend in the SA cruise data, in contrast to Cape Point where *Slemr et al.* [2011] reported a 1996-2010 decreasing trend of $-0.034 \pm 0.005 \text{ ng m}^{-3} \text{ a}^{-1}$, corroborated by their five SA ship cruises. However, Figure 1 shows that these five cruises (filled squares) may not be representative of the ensemble of cruise data in the SA. In particular, the low 2008-2009 concentrations from *Kuss et al.* [2011] that anchor the *Slemr et al.* [2011] decreasing trend are not consistent with three other 2006-2010 cruises that show higher values.

We do not have a good explanation for the inconsistency in 1996-2010 trends between the SA cruises and Cape Point. However, atmospheric Hg at Cape Point exhibits anomalous behavior. Frequent Hg depletion events have been recorded at that site since the introduction of high-resolution measurements in 2007 and appear to be of local origin (indicated by low wind speed) but are otherwise not understood [*Brunke et al.*, 2010]. Of the eight SA cruises that used high-resolution Tekran measurements, only the *Kuss et al.* [2011] cruise reported evidence of sporadic depletion events [*Brunke et al.*, 2010]. Interpretation of the Cape Point trend may require better understanding of factors determining local Hg depletion events.

Slemr et al. [2011] argued that the 1996-2009 decrease in Hg was worldwide, mainly on the basis of the similarity between Mace Head, Cape Point, and their selected cruises. We find otherwise from the ensemble of cruise data, with a sharp contrast between the strong decreasing trend in the NA and the lack of trend in the SA. Such a hemispheric contrast is not inconsistent

with the atmospheric lifetime of ~6 months for Hg that allows it to mix efficiently on a hemispheric scale but less so between hemispheres [Corbitt *et al.*, 2011]. It appears furthermore that the decrease over the NA is stronger than observed elsewhere in the Northern hemisphere (NH) over the past 20 years, including at the nearby continental sites of Mace Head, Kuujuaupik, and St. Anicet (-0.024 to -0.033 ng m³ a⁻¹ [Ebinghaus *et al.*, 2011; Slemr *et al.*, 2011; Cole *et al.*, 2011]) and at Arctic sites (-0.009 ng m⁻³ a⁻¹ at Alert and no trend at Svalbard) [Cole and Steffen, 2010; Cole *et al.*, 2011]).

Concentrations of Hg in the MBL reach steady state with the surface ocean mixed layer on a time scale of less than a year [Soerensen *et al.*, 2010b]. Surface seawater Hg⁰ data were collected during five cruises in the Northwest Atlantic in 1998-2000 [Mason *et al.*, 2001] and five cruises in the same area in 2008-2010 [Andersson *et al.*, 2008; Mason *et al.*, 2009; R.P. Mason, unpublished data]. The data are given in Auxiliary Table A2. Observed mean concentrations were 244±100 fM in 1998-2000 and 136±33 fM in 2008-2010. This represents a significant decrease of -5.7% a⁻¹ during the period (t-test paired means difference, $p < 0.05$). Thus, surface water Hg⁰ data support the decline found in the NA surface air and are even steeper than in surface air over the same 1999-2009 period (-0.066 ng m³ a⁻¹, -4.0% a⁻¹, $n = 14$). As we will see below, this is consistent with oceanic forcing of the atmospheric trend.

3. Cause of the North Atlantic Hg decline

We use the GEOS-Chem global biogeochemical mercury model (v 9-01-02; <http://www.geos-chem.org>) to investigate possible causes of the Hg trend in the NA. The model is described by Soerensen *et al.* [2010b], Holmes *et al.* [2010], and Zhang *et al.* [2012] and has been extensively evaluated against observed concentrations and wet deposition from surface sites and ship cruises. The model couples a 3-D atmosphere with 2-D (horizontal) surface ocean and land

reservoirs. Subsurface ocean concentrations are specified as constant values from mean observed concentrations in different ocean basins (Soerensen et al., 2010b). Here ‘surface ocean’ denotes the ocean mixed layer (annual global mean depth ~50 m), and ‘subsurface ocean’ denotes the water column from the bottom of the mixed layer to the depth of the permanent thermocline (~1000-1500 m)[Mason et al., 2012]. We updated the standard simulation of aqueous-particle partitioning of Hg^{II} in the surface ocean [Soerensen et al., 2010b] with a specific K_D (affinity of aqueous Hg^{II} for the solid phase) of $1 \times 10^5 \text{ L kg}^{-1}$ for the NA based on observed values from Mason et al. [1998]. All simulations presented here use $4^\circ \times 5^\circ$ horizontal resolution, are initialized for two years (2006-2007) and use the same meteorological year (2008).

Global anthropogenic Hg emissions increased by 30% between 1990 and 2008 according to Streets et al. [2011] but Hg^{II} emissions in the USA and Europe decreased by -20% and -40%, respectively. Hg^{II} is removed regionally by deposition, in contrast to Hg⁰, which mixes on a hemispheric scale. Thus, decreases in locally emitted Hg^{II} could have selectively impacted the North Atlantic seawater Hg levels and associated evasion rates. We conducted GEOS-Chem simulations using 1990, 2000, and 2008 anthropogenic emissions from Streets et al. [2011] with all else kept constant. As shown in Figure 2, regional decrease of Hg^{II} emissions is insufficient to compensate for the global rise of Hg emissions except over western Eurasia. GEOS-Chem includes fast in-plume reduction of Hg^{II} from power plant emissions [Zang et al., 2012], but a sensitivity simulation without this in-plume reduction still does not show a 1990-2008 decline over the NA.

In the 1980s and 1990s NA subsurface waters were found to be enriched in total Hg compared to the North Pacific [Gill and Fitzgerald, 1988; Laurier et al., 2004]. However there is evidence that North Atlantic subsurface seawater concentrations have declined sharply over the

past several decades. Profile measurements near the Bermuda Atlantic Time-Series Study (BATS) station (32N, 64W) showed values of 5-9 pM in July 1979 and September 1983 [Gill and Fitzgerald, 1988], 1.0-2.5 pM in different seasons in 1999-2000 [Mason et al., 2001], and 0.6-0.8 pM in June 2008 [Mason et al., 2012]. Similarly, Scotian Shelf inflow water measured by Dalziel [1992] in the mid-1980s showed more than a two-fold enrichment compared to concentrations measured in 2002 [Sunderland et al., 2012]. Clean measurement techniques were established in the 1980s [Gill and Fitzgerald, 1987] and allow confidence in these trends.

We conducted GEOS-Chem simulations imposing different subsurface ocean concentrations in the NA (0-70°N) for 1990, 2000 and 2008, with all else constant including anthropogenic emissions at 2008 levels (Streets et al. 2011). Since the surface air data show no significant trend for 1977-1990 (Figure 1), we assume a subsurface ocean concentration of 5 pM (lower range from Gill and Fitzgerald [1988]) in 1990, 1.8 pM in 2000, and 0.7 pM in 2008. Modeled Hg concentrations in the surface ocean and atmosphere respond to changes in subsurface ocean Hg within 5-6 months [Corbitt et al., 2011], so that individual simulations for each decade can adequately capture the 1990-2008 trend.

Figure 2 shows that changes in NA subsurface seawater concentrations results in a $-0.042 \text{ ng m}^{-3} \text{ a}^{-1}$ (-2.4%) 1990-2008 model decline in surface air concentrations across the NA, closely reproducing the observed trend of $-0.046 \pm 0.010 \text{ ng m}^{-3} \text{ a}^{-1}$ (Figure 1). The simulated surface ocean Hg⁰ concentration in the NA declines by $-7.0 \% \text{ a}^{-1}$ between 2000 and 2008, close to the observed $-5.7 \% \text{ a}^{-1}$ decline. The steeper decline in the NA surface ocean relative to the atmosphere is thus reproduced by the model, providing further support for a subsurface ocean driver of the atmospheric trend.

Simulations show that declining Hg concentrations in subsurface seawater have a spatially variable impact on surface air Hg⁰ concentrations, with more pronounced effects on the NA MBL and Coastal Europe surface air Hg⁰ compared to elsewhere in the NH (Figure 2). Using the 1990-2008 model estimates we capture the yearly percentage decline of Hg⁰ at mid latitude terrestrial sites within an average of $\pm 28\%$ [Ebinghaus *et al.*, 2011; Cole *et al.*, 2011]. Thus, declines in subsurface seawater Hg concentrations help to explain variable rates of declines in the surface atmosphere observed across sites in the NH. The simulated trend in South Atlantic surface air is $-0.8\% \text{ a}^{-1}$ for the period 1990 to 2008 (Figure 1). This is much weaker than in the Northern Hemisphere as would be expected from the atmospheric lifetime of Hg⁰ [Corbitt *et al.*, 2011]. A simple bootstrapping procedure shows that given the variability in cruise observations the weak modeled trend over the South Atlantic would be undetectable in observations (Figure 2).

4. Drivers of the North Atlantic subsurface water decline

Vertical seawater profiles from the West Atlantic [Gill and Fitzgerald, 1988; Mason *et al.*, 2001, 2012] suggest an 80% decline in NA subsurface waters in the last 30 years. A parallel can be drawn with Pb concentrations, which declined by 85% in the subsurface NA between 1979 and 2008 after leaded gasoline was phased out in the 1970s and 1980s [Wu and Boyle, 1997; Lee *et al.*, 2011]. Tracer data for NA subsurface waters indicate a mean age of 10-30 years since last contact with the atmosphere [Fine *et al.*, 2010]. Thus a 30-year trend in the Hg content of this reservoir would be sensitive to trends in inputs going back to the 1960s. We consider different possibilities below.

Parrella *et al.* [2012] suggested that decreased Hg input to NA subsurface waters over the past decades could have been caused by a decrease in Hg⁰ oxidation in the MBL, which

appears to be the principal supplier for Hg deposition to the NA [Selin *et al.*, 2008; Holmes *et al.*, 2009]. If Br atoms are the main oxidant for Hg⁰ in the MBL [Hedgecock and Pirrone, 2004; Holmes *et al.*, 2009], then a large decrease in Hg⁰ oxidation would be expected because of the observed two to three times increase of surface ozone over the NA over the past five decades [Lelieveld *et al.*, 2004; Parrish *et al.*, 2009]. Increasing surface ozone causes a proportional decrease in Br concentrations by shifting the Br/BrO radical equilibrium toward BrO.

Large declines in Hg inputs from rivers and wastewater to the NA Ocean concurrent with regulation of industry and other point sources are another possible driver of the temporal trend in subsurface NA seawater between the 1970s and present. Rivers are presently estimated to provide up to 220 Mg a⁻¹ to the NA Ocean compared to ~860 Mg a⁻¹ from atmospheric deposition [Sunderland and Mason, 2007]. However, a decline of ~50% in this source relative to peak discharges in the 1970s is observed in multiple sedimentary archives collected at the mouths of rivers from estuaries in North America and Europe [e.g., Sunderland *et al.* 2010; Varekamp *et al.*, 2000; Boutier *et al.*, 2011; Elbaz-Poulichet *et al.*, 2012] suggesting a much larger historical supply of Hg to the NA.

Chlor-alkali plants were a particularly large (losses > 875 Mg a⁻¹) Hg source in the 1970s in the US and Europe [D'Itri and D'Itri, 1977, Euro Chlor, 2005]. The European chlor-alkali industry reported that Hg discharged in effluents almost equaled atmospheric Hg losses in the 1970s [Euro Chlor, 2005]. Many chlor-alkali factories were located in coastal areas with effluents containing Hg discharged directly into estuaries [D'Itri and D'Itri, 1977]. The Hg source from the chlor-alkali industry is now only a few percent of what it was in the 1970s [D'Itri and D'Itri, 1977, Euro Chlor, 2005; Chlorine Institute, 2009].

Municipal wastewater discharged directly into marine ecosystems is another Hg source that has declined dramatically over the past several decades. For example, measured Hg in wastewater from the largest facility serving the city of Boston indicates that Hg concentrations declined by >95% between 1991 and 2009 due to implementation of secondary wastewater treatment [NOAA, 1994; Wu, 2011]. Phase-out of Hg from consumer products and widespread introduction of secondary municipal wastewater treatment both likely caused large declines in Hg inputs to ocean margins.

We thus hypothesize that wastewater and riverine Hg were major historical sources of Hg to the NA ocean and subsequently to the atmosphere. We suggest that phasing out of Hg in wastewater and many industries in North America and Europe over the past decades is at least partly responsible for the rapid decline of atmospheric concentrations over the NA and nearby continents. Timescales of Hg cycling in subsurface waters of the NA means that presently much of the historic Hg has accumulated in the deep ocean or dispersed globally through evasion. Thus, we expect the decline of NH atmospheric Hg concentrations to slow down, or possibly reverse if global emissions continue to rise. Our work highlights the need for better understanding the contribution of effluent Hg releases to global anthropogenic Hg both in the past and in the present.

Acknowledgements

We acknowledge financial support for this work from NSF Atmospheric Chemistry, NSF Chemical Oceanography, and the Electric Power Research Institute (EPRI).

References

Andersson, M.E., W.F. Fitzgerald, and R.P. Mason(2008), Mercury air-sea exchange on the New England shelf, paper presented at AGU fall meeting 2008, AGU, San Francisco, USA.

Amos, H., D.J. Jacob, C.D. Holmes, J.A. Fisher, Q. Wang, R.M. Yantosca, E.S. Corbitt, E. Galarneau, A.P. Rutter, M.S. Gustin, A. Steffen, J.J. Schauer, J.A. Graydon, V.L.st. Louis, R.W. Tablbot, E.S. Edgerton, Y. Zhang, and E.M. Sunderland (2011), Gas-particle partitioning of atmospheric Hg(II) and its effect on global mercury deposition, *Atmos. Chem. Phys.*, *12*, 591-603, doi: 10.5194/acp-12-591-2012.

Aspmo, K., C. Temme, T. Berg, C. Ferrari, P. A. Gauchard, X. Fain, and G. Wibetoe (2006), Mercury in the atmosphere, snow and melt water ponds in the North Atlantic Ocean during Arctic summer, *Environ. Sci. Technol.*,*40*(13), 4083-4089.

Boutier, B., J. Quintin, E. Rozuel, A. Dominique, and J. Bretaudeau-Sanjuan (2011), Retrospective study of metal contamination time trends in the French part of the Bay of Biscay, *Environ. Technol.*, *32*(15), 1807-1815.

Brunke, E. G., C. Labuschagne, R. Ebinghaus, H. H. Kock and F. Slemr (2010), Gaseous elemental mercury depletion events observed at Cape Point during 2007-2008, *Atmos.Chem. Phys.*,*10*(3), 1121-1131.

Chlorine Institute (2009), Twelfth annual report to EPA, available at www.epa.gov/Region5/mercury/pdfs/12thcl2report.pdf (accessed March 2012).

Cole, A.S., and A. Steffen (2010), Trends in long-term gaseous mercury observations in the Arctic and effects of temperature and other atmospheric conditions, *Atmos. Chem. Phys.*, *10*, 4661-4672, doi:10.5194/acp-10-4661-2010.

Cole, A.S., A. Steffen, L. Poissant, M. Pilote, K.A. Pfaffhuber, and T. Berg(2011), Trends in atmospheric mercury concentrations in the Northern Hemisphere: Why is the Arctic different?

Paper presented at the 10th International Conference on Mercury as a Global Pollutant, Halifax, Canada.

Corbitt, E.S., D.J. Jacob, C.D. Holmes, D.G. Streets, and E.M. Sunderland(2011), Global Source-Receptor Relationships for Mercury Deposition Under Present-Day and 2050 Emissions Scenarios, *Environ. Sci.Technol.*, *45*, 10477-10484.

Dalziel, J.A. (1992), Reactive mercury on the Scotian Shelf and in the adjacent northwest Atlantic Ocean, *Mar. Chem.*, *37*, 171-178.

D'Itri, P.A., and F.M. D'Itri (1977), *Mercury contamination: A human tragedy*, John Wiley & Sons, Inc.: USA, pp. 139-161.

Ebinghaus, R., S.G. Jennings, H.H. Kock, R.G. Derwant, A.J. Manning, and T.G. Spain(2011), Decreasing trends in total gaseous mercury observations in baseline air at Mace Head, Ireland from 1996 to 2009, *Atmos. Environ.*, *45*, 3475-3480.

Elbaz-Poulichet, F, L. Dezileau, R. Freydier, D. Cossa, and P. Sabatier (2012), A 3500-year record of Hg and Pb contamination in a Mediterranean sedimentary archive (The Pierre Blanche Lagoon, France), *Environ. Sci. Technol.*, *45*, 8642-8647.

Euro Chlor (2005), The European Chlor-alkali Industry, Progress Report 2005, available at <http://eurochlor.clients.cwndesign.co.uk/upload/documents/document161.pdf> (accessed April 2012).

Fine, R.A. (2010), CFCs in the Ocean, in *Marine Chemistry and Geochemistry*, edited by K.K. Turekian, pp. 155-162, Academic Press, San Diego, California.

Fischer, E.V., D.A. Jaffe, and E.C. Weatherhead (2011), Free tropospheric peroxyacetyl nitrate (PAN) and ozone at Mount Bchelor: potential causes of variability and timescale for trend detection, *Atmos. Chem. Phys.*, *11*, 5641-5654, doi: 10.5194/acp-11-5641-2011.

Gill, G.A., and W.F. Fitzgerald (1987), Picomolar mercury measurements in seawater and other materials using stannous chloride reduction and two-stage gold amalgamation with gas phase detection, *Mar. Chem.*, *20*, 227-243.

Gill, G. A. and W. F. Fitzgerald, Vertical Mercury Distributions in the Oceans, *Geochim. Cosmochim. Acta*,*52(6)*, 1719-1728, 1988.

Gustin, M. and D. Jaffe (2010), Reducing the Uncertainty in Measurement and Understanding of Mercury in the Atmosphere, *Environ. Sci. Technol.*,*44(7)*, 2222-2227.

Hedgecock, I. M. and N. Pirrone (2004), Chasing quicksilver: Modeling the atmospheric lifetime of Hg⁰(g) in the marine boundary layer at various latitudes, *Environ. Sci. Technol.*, *38 (1)*, 69-76.

Holmes, C. D., D. J. Jacob, R. P. Mason and D. A. Jaffe (2009), Sources and deposition of reactive gaseous mercury in the marine atmosphere, *Atmos. Environ*, *43 (14)*, 2278-2285.

Holmes, C. D., D. J. Jacob, E. S. Corbitt, J. Mao, X. Yang, R. Talbot and F. Slemr (2010), Global atmospheric model for mercury including oxidation by bromine atoms, *Atmos. Chem. Phys.*,*10(24)*, 12037-12057.

Kuss, J., C. Zulficke, C. Pohl, and B. Schneider (2011), Atlantic mercury emission determined from continuous analysis of the elemental mercury sea-air concentration difference within transects between 50°N and 50°S, *Global Biogeochem. Cycles*, *25*, GB3021, doi:10.1029/2010GB003998.

Lamborg, C. H., K. R. Rolfhus, W. F. Fitzgerald and G. Kim (1999), The atmospheric cycling and air-sea exchange of mercury species in the South and equatorial Atlantic Ocean, *Deep-Sea Res.*,*46(5)*, 957-977.

Laurier, F. J. G., R. P. Mason, G. A. Gill and L. Whalin (2004), Mercury distributions in the North Pacific Ocean - 20 years of observations, *Mar. Chem.*,*90(1-4)*, 3-19.

Laurier, F. and R. Mason (2007), Mercury concentration and speciation in the coastal and open ocean boundary layer, *J. Geophys. Res. [Atmos]*, *112(D6)*, D06302, doi:10.1029/2006JD007320.

Lee, J-M., E.A. Boyle, Y. Echevoyen-Sanz, J.N. Fitzsimmons, R. Zhang, and R.A. Kayser (2011), Analysis of trace metals (Cu, Cd, Pb, and Fe) in seawater using single batch nitrilotriacetate resin extraction and isotope dilution inductively coupled plasma mass spectrometry, *Analyt. Chim. Acta*, *686*, 93-101, doi: 10.1016/j.aca.2010.11.052.

Lelieveld, J., J. van Aardenne, H. Fischer, M. de Reus, J. Williams, and P. Winkler (2004), Increasing Ozone over the Atlantic Ocean (2004), *Science*, *304*, 1483-1487, doi: 10.1126/science.1096777.

Mahaffey, K. R., R. P. Clickner and R. A. Jeffries (2009), Adult Women's Blood Mercury Concentrations Vary Regionally in the United States: Association with Patterns of Fish Consumption (NHANES 1999-2004), *Environ. Health Perspec.*, *117 (1)*, 47-53.

Mason, R. P., K. R. Rolfhus and W. F. Fitzgerald (1998), Mercury in the North Atlantic, *Mar. Chem.*, *61(1-2)*, 37-53.

Mason, R. P., N. M. Lawson and G. R. Sheu (2001), Mercury in the Atlantic Ocean: factors controlling air-sea exchange of mercury and its distribution in the upper waters, *Deep-Sea Res. II*, *48(13)*, 2829-2853.

Mason, R.P., M.E. Andersson, A.L. Soerensen, and E.M. Sunderland (2009), Measurements and modeling of the air-sea exchange of mercury, paper presented at AGU fall meeting 2009, American Geophysical Union, San Fransisco, USA.

Mason, R.P., A.L. Choi, W.F. Fitzgerald, C.R. Hammerschmidt, C.H. Lamborg, A.L. Soerensen, E.M. Sunderland(2012), Mercury biogeochemical cycling in the ocean and policy implication, *Environ.Res.*, in press.

Mergler, D., H. A. Anderson, L. H. M. Chan, K. R. Mahaffey, M. Murray, M. Sakamoto and A. H. Stern (2007), Methylmercury exposure and health effects in humans: A worldwide concern, *Ambio*, 36(1), 3-11.

NOAA (1994), Gulf of Maine Point Source Inventory: A Summary by Watershed for 1991. Pollution Characterization Branch, Strategic Environmental Assessments Division, National Oceanic and Atmospheric Administration. 164 pp., Silver Spring, MD.

Parrella, J.P., D.J. Jacob, Q. Liang, Y. Zhang, L.J. Mickley, B. Miller, M.J. Evans, X. Yang, J.A. Pyle, N. Theys, and M. Van Roozendaal(2012), Tropospheric bromine chemistry: implications for present and pre-industrial ozone and mercury, *Atmos. Chem. Phys. Discuss.*, 12, 9665-9715, doi:10.5194/acpd-12-9665-2012.

Parrish, D.D., D.B. Miller, and A.H. Goldstein(2009), Increasing ozone in marine boundary layer inflow at the west coast of North America and Europe, *Atmos. Chem. Phys.*, 9, 1303-1323.

Selin, N. E., D. J. Jacob, R.M. Yantosca, S. Strode, L. Jaegle, and E.M. Sunderland (2008), Global 3-D land-ocean-atmosphere model for mercury: Present-day versus preindustrial cycles and anthropogenic enrichment factors for deposition, *Glob. Biochem. Cycles*, 22, GB2011.

Slemr, F., E.-G. Brunke, R. Ebinghaus, and J. Kuss (2011), Worldwide trend of atmospheric mercury since 1995, *Atmos. Chem. Phys.*, 11, 4779-4787, doi:10.5194/acp-11-4779-2011.

Soerensen, A.L., H. Skov, D.J. Jacob, B.T. Soerensen, and M.S. Johnson(2010a), Global concentrations of gaseous elemental mercury and reactive gaseous mercury in the marine boundary layer, *Environ. Sci. Technol.*, 44, 7425-7430, doi:10.1021/es903839n.

Soerensen, A.L., E.M. Sunderland, C.D. Holmes, D.J. Jacob, R.M. Yantosca, H. Skov, J.H. Christensen, S.H. Strode, and R.P. Mason(2010b), An improved global model for air-sea

exchange of mercury: High concentrations over the North Atlantic, *Environ. Sci. Technol.*, *44*, 8574-8580, doi:10.1021/es102032g.

Sommar, J., M. E. Andersson and H. W. Jacobi (2010), Circumpolar measurements of speciated mercury, ozone and carbon monoxide in the boundary layer of the Arctic Ocean, *Atmos. Chem. Phys.*, *10*(11), 5031-5045.

Streets, D.G., M.K. Devane, Z. Lu, T.C. Bond, E.M. Sunderland, and D.J. Jacob (2011), All-Time Release of Mercury to the Atmosphere from Human Activities, *Environ. Sci. Technol.*, *45*(24), 10485-10491.

Sunderland, E. M. and R. P. Mason (2007), Human impacts on open ocean mercury concentrations, *Global Biogeochem. Cycles*, *21*(4).

Sunderland, E.M., J. Dalziel, A. Heyes, B.A. Branfireun, D.P. Krabbenhoft, and F.A.P.C.

Gobas(2010), Response of a Macrotidal Estuary to Changes in Anthropogenic Mercury Loading between 1850 and 2000, *Environ. Sci. Technol.*, *44*, 1698-1704.

Sunderland, E.M., A. Amirbahman, N.M. Burges, J. Dalziel, G. Harding, S.H. Jones, E. Kamai, M.R. Karagas, X. Shi, and C.Y. Chen(2012), Mercury Sources and Fate in the Gulf of Maine, in press.

Temme, C., F. Slemr, R. Ebinghaus and J. W. Einax (2003), Distribution of mercury over the Atlantic Ocean in 1996 and 1999-2001, *Atmos. Environ.*, *37*(14), 1889-1897.

Temme, C., P. Blanchard, A. Steffen, C. Banic, S. Beauchamp, L. Poissant, R. Tordon and B. Wiens (2007), Trend, seasonal and multivariate analysis study of total gaseous mercury data from the Canadian atmospheric mercury measurement network (CAMNet), *Atmos. Environ.*, *41*(26), 5423-5441.

Toerseth, K., W. Aas, K. Breivik, A.M. Fjaeraa, M. Fiebig, A.G. Hjellbrekke, C. Lund Myhre, S. Solberg, and K.E. Yttri(2012), Introduction to the European Monitoring and Evaluation Programme (EMEP) and observed atmospheric composition change during 1972-2009, *Atmos. Chem. Phys. Discuss.*, *12*, 1733-1820, doi:10.5194/acpd-12-1733-2012.

Varekamp J.C., M.R. Buchholtz ten Brink, E.L. Mccray, and B. Kreulen (2000), Mercury in Long Island Sound Sediments, *J. Coast. Res.*, *16*(3), 613-626.

Wu, D. (2011), NPDES compliance summary report, fiscal year 2009. Massachusetts Water Resources Authority, Environmental Quality Department. Report ENQUAD 2011-04. 135 pp., Boston, Massachusetts.

Wu, J., and E.A. Boyle (1997), Lead in the western North Atlantic Ocean: Completed response to leaded gasoline phaseout, *Geochim. Cosmochim. Acta*, *61*(15), 3279-3283.

Zhang, Y., L. Jaegle, A. von Donkelaar, R.V. Martin, C.D. Holmes, H.A. Amos, Q. Wang, R. Talbot, R. Artz, S. Brooks, W. Luke, T.M. Holsen, D. Felton, E.K. Miller, K.D. Perry, D. Schmeltz, A. Steffen, R. Tordon, P. Weiss-Penzias, and R. Zsolway (2012), Nested-grid simulation of mercury over North America, *Atmos. Chem. Phys. Discuss.*, *12*,2603-2646, doi: 10.5194/acpd-12-2603-2012.

Figure Captions

Figure 1. Ship cruises used in the surface air Hg trend analysis (left). Cruise data are from *Temme et al.* [2003] (1, 2, 3, 4, 5, 7, 9, 12, 13); *Mason et al.* [1998] (6); *Lamborg et al.* [1999] (8); *Mason et al.* [2001] (10,11,14); *Laurier and Mason* [2007] (15); *Aspmo et al.* [2006] (16); *Sommar et al.* [2010] (17); *Soerensen et al.* [2010a] (18, 19, 20); *Kuss et al.* [2011] (21, 24), *Andersson et al.* [2008] (22), *Mason et al.* [2009] (23, 25), M. Witt (manuscript in preparation,

2012) (26), R. Ebinghaus (unpublished data, 2010) (27). Hg concentration trends in surface air of the NA (upper right) and SA (lower right), 1977-2009, reported as total gaseous mercury (TGM) or elemental mercury (Hg^0) which we view here as equivalent (see text). Squares represent mean concentrations measured on ship cruises, with vertical bars indicating standard deviations. The regression line for the observed 1990-2009 trend in NA is shown ($0.046 \pm 0.010 \text{ ng m}^{-3} \text{ a}^{-1}$, $p < 0.001$, $n = 18$). Model trends for NA including changes in anthropogenic emissions or in subsurface NA are also shown. Red circles are yearly medians for Mace Head (Ireland) and Cape Point (South Africa) taken from *Slemr et al.* [2011]. Filled squares represent the cruise observations used by *Slemr et al.* [2011] in their trend analysis. Further information on cruises is found in the Auxiliary Table A1.

Figure 2. Simulated differences of Hg surface air concentrations in GEOS-Chem between 1990 and 2008 as driven by changes in anthropogenic emissions (left) and subsurface ocean concentrations of the NA for 0° - 70°N (right). The results are expressed as an annual trend (a^{-1}).