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1	Levels and distributions of Dechlorane Plus
2	in coastal sediment of North China
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12	
13	ABSTRACT
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15	Dechlorane Plus (DP) has been determined in surface sediments from three Chinese coastal
16	bays, e.g. Jiaozhou, Sishili and Taozi Bay in North China. DP concentrations ranged from
17	<1.2 to 187pg/g dry weight (dw) (mean: 24.7 pg/g dw) in Jiaozhou Bay, <1.2 to 135 pg/g dw
18	(mean 69.9 pg/g dw) in Sishili Bay and <1.2 to 66.7 pg/g dw (mean: 40.4 pg/g dw) in Taozi
19	Bay, respectively. Additionally, two dechlorinated species were quantified, which accounted
20	for 0.6% to 5.1% in the \sum DP concentration.
21	The f_{syn} values (syn-isomer/(syn- + anti- isomer)) in sediments from Jiaozhou Bay (mean
22	0.29) were close to technical DP mixture (0.2 to 0.4), probably indicating a local usage of DP.

23 In contrast, sediments in Sishili and Taozi Bay showed much lower f_{syn} values (mean

0.16).During transportation the DP isomers underwent stereo selective degradation which
partly resulted in the enrichment of *anti*-DP in coastal sediment.

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27 Keywords: Dechlorane Plus, coastal sediment, North China

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Capsule: study for occurrence of Dechlorane Plus in sediment of coastal environment ofNorth China

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32 **1. Introduction**

33 Dechlorane Plus (DP, $C_{18}H_{12}Cl_{12}$), the common name of bis(hexachlorocyclopentadieno) 34 cyclooctane, which is used as a substitute for the toxic Declorane (Mirex) has been produced 35 for about four decades (Gauthier et al., 2007; Qiu et al., 2007). As a highly chlorinated flame 36 retardant, DP has been used in plastic roofing material, hardware connectors for computers 37 and for coating electrical wires and cables (Tomy et al., 2008). The annual production of DP 38 was reported as 5000 tons, and it is sold in North American, Europe and Asia (Kang et al., 39 2010; Wang et al., 2010a). As a result of the wide application, DP has been detected in 40 various environmental compartments, such as air, water, sediment, fish and serum of human 41 being (Hoh et al., 2006; Tomy et al., 2007; Ren et al., 2009; Qi et al., 2010; Wang et al., 42 2010b). Bioaccumulation and biomagnification were reported for organisms of high tropic 43 levels, which suggest a potential threat for humans (Tomy et al., 2008; Ren et al., 2009).

In China DP has been produced by Anpon Corporation for seven years, with a total volume of 2100 to 7000 tons. This manufacturing plant has been recognized as the main DP source for the surrounding area (Wang et al., 2010b). The DP concentration in soil near the plant was 1490 ± 3580 ng/g dry weight (dw), which is one magnitude higher than the maximum concentration in Lake Ontario (586 ng/g dw) (Sverko et al., 2008; Wang et al., 2010b). The usage and unintended disposal spread DP in non-producing areas. Ren et al. (2008) collected air samples from 97 Chinese urban and rural sites, and stated that the highest concentration was found in Kunming, a famous tourist city far away from the DP manufactory. In Harbin, an industrial city of Northeast China without DP manufacture, the mean concentration of DP was 0.11 ± 0.05 ng/g dw in urban sediments (Qi et al., 2010). Besides domestic usage, the import of DP or DP containing products (including e-waste) from overseas could also elevate the levels in the Chinese environment. Yu et al. (2010) found that in an e-waste recycling site in Qingyuan, the concentration of DP reached up to 3327 ng/g dw in surface soil.

57 Sediments have been regarded as one of the major sinks for persistent organic pollutants (POPs), e.g., polychlorinated biphenyls (PCBs), organochlorine pesticides (OCPs) and 58 59 brominated flame retardants (BFRs) (Hung et al., 2006; Minh et al., 2007; Guzzella et al., 2008). These pollutants enter the marine sediments through atmospheric deposition, riverine 60 61 runoff and direct release from human activities on the ocean (Moon et al., 2007b). Both, the 62 syn- and anti- isomers have a high octanol-water partition coefficient ($\log K_{ow}$) value of 9.3, which is similar to some polybrominated diphenyl ethers (PBDEs) that strongly adsorbed to 63 64 sediments (Palm et al., 2002; Moon et al., 2007a; Sverko et al., 2008). The OxyChem report 65 employing EPIWIN version 3.12 presented that 66.6% of the DP was distributed in sediments (US, 2008). High concentrations of DP have been detected in sediments from Lake Ontario 66 67 with an inventory of DP calculated to about 20 tons (Qiu et al., 2007; Sverko et al., 2008). 68 Presently, only few data are available on the distribution of DP in coastal sediments. Our 69 present study focuses on the concentrations, distributions and possible sources of DP in 70 surface sediments of three Chinese bays.

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- 72 2. Materials and Methods
- 73 2.1 Sample Collection

Surface sediments (0-5 cm) were collected in coastal zones of Yellow Sea in April, 2010.
The locations of all the 48 samples are shown in Fig. 1. Twenty-five samples were collected

in the Jiaozhou Bay which is surrounded by Qingdao City (35°35'-37°09'N, 119°30'-76 77 121°00'E), and four river sediment samples were collected in Dagu River and Qingdao River, discharging into the Jiaozhou Bay. Six samples were collected in Taozi Bay and thirteen in 78 Sishili Bay near Yantai City (36°16'-38°23'N, 119°34'-121°57'E). The sediments were 79 80 collected by stainless spades and then sealed in polyethylene (PE) bags. They were frozen 81 immediately after sampling at -20 °C until extraction. Prior to extraction, sediments were 82 freeze-dried for three days at -55°C and then ground, homogenized by agate mortar and pestle 83 and wrapped up in pre-cleaned filter papers.

84 2.2 Extraction and cleanup

85 All solvents were analytical grade and were re-distilled before using. Filter papers were Soxhlet extracted for 72 h using dichloromethane (DCM). Samples were spiked with 20 ng of 86 87 PCB 209 as surrogate standard prior to extraction. 20 g of sediments were extracted in a 88 Soxhlet apparatus for 72 h using DCM. Activated copper slices were added to the collection 89 flask to remove elemental sulfur. Extracts were evaporated to 10 mL and the solvent was changed to hexane before further evaporation to <2 mL. A 8 mm i.d. modified column packed 90 91 with 1 cm anhydrous sodium sulfate, 3 cm 50% sulfuric acid silica, 3 cm neutral silica gel and 92 3 cm neutral alumina from the top to the bottom was employed for clean-up. The silica was 93 precleaned with acetone and DCM, activated at 180 °C for 12 h and further 3 % deactivated 94 using Millipore water. The neutral alumina was treated in a similar way, activated at 250 °C. 95 Sulfuric acid silica was prepared by adding 50 % (w/w) HPLC grade sulfuric acid (98 %) to 96 the deactivated silica. The anhydrous sodium sulfate was baked at 450 °C for 4h. The fraction 97 was concentrated to 50 µL under a gentle high-purity nitrogen (>99.99 %) stream. 20 ng of 98 BDE-77 was spiked as an internal standards before injection.

99 2.3 Instrument Analysis

100 The standards of *syn*-DP, *anti*-DP, and two dechlorinated species, $aCl_{11}DP$ and $aCl_{10}DP$, 101 were obtained from Wellington Laboratories. An Agilent 7890A GC equipped with a 30 m ×

102 0.25 mm i.d. (0.25 µm film thickness, J&W Scientific) DB-5 fused silica capillary column 103 was connected to a 5975C MSD under a negative chemical ionization (NCI) mode using 104 methane as the reagent gas. The injector temperature was 280 °C employed in splitless mode 105 with 1 µL injection volume. The oven program was as follows: initial 60 °C for 2 min, 30 °C min⁻¹ to 180 °C, 2 °C min⁻¹ to 280 °C, 30 °C min⁻¹ to 300 °C and held for 6 min, and then 106 ramped at 30 °C min⁻¹ to 310 °C and held for a final 7 min. MS was operated in single ion 107 monitoring (SIM) mode with ion source, quadrupole and transfer line temperatures held at 108 109 150, 150 and 280 °C, respectively. The following ions were monitored: m/z 653.8, 617.9 and 110 583.9 for syn- and anti-DP; m/z 617.8, 583.7 and 547.8 for aCl₁₁DP; m/z 583.8, 549.9 and 111 513.8 for aCl₁₀DP; *m/z* 79.0 and 81.0 for BDE77; and *m/z* 497.7 and 499.7 for PCB209.

112 *2.4 QA/QC*

113 The criteria for the identification and quantification of target compounds are given as 114 follow: (1) The retention times matched those of the standard compounds within ± 0.05 min. 115 (2) The signal-to-noise (S/N) ratio of all peaks was greater than 5:1. (3) The theoretical 116 isotopic ratios of the qualifier ions were within \pm 15% of the standard values. The linear 117 dynamic range of the instrument was between 2 to 25 pg on the column ($R^2 > 0.996$) for DP 118 isomers and their dechlorinated species. PCB 209 was spiked in every sample as DP recovery 119 indicator. The mean recovery rate was 96±12%. Concentrations of sediments were not 120 recovery corrected. In three procedure blanks, both DP isomers and their dechlorinated 121 moieties were not detected. Method detection limits (MDLs) were calculated based on the 122 instrumental S/N ratios of 10. They were 0.4 pg/g for syn-DP, 0.8 pg/g for anti-DP, 0.5 pg/g 123 for aCl₁₀DP and 0.9 pg/g for aCl₁₁DP, respectively.

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125 **3. Results and Discussion**

126 3.1 Concentrations and spatial distribution of Dechlorane Plus

127 The concentrations of total DP (syn-+anti-DP) in river and marine sediments are shown in 128 Table 1. DP was detected in all surface sediment samples from Jiaozhou, Taozi and Sishili 129 Bay. Large spatial variations of total DP were observed, with <1.2 to 187 pg/g in Jiaozhou 130 Bay, <1.2 to 66.7 pg/g in Taozi Bay and <1.2 to 135 pg/g in Sishili Bay, respectively. Qi et al. 131 (2010) reported DP in sediments from Songhua River, Northeast China. The concentrations (> 132 4.5 to 160 pg/g dw) were similar to those found in this study. Comparing with data from the 133 sediments near manufactory locations, the concentrations in the three bays were one to three 134 orders of magnitude lower. Wang et al. (2010) reported concentrations ranging from 1.86 to 135 8.0 ng/g dw in the canal sediments near Anpon Corporation in Huai'an, China, and Sverko et 136 al. (2008) reported high concentrations in sediment of Great Lakes ranging from 0.061 to 586 137 ng/g near the OxyChem company. According to Wang et al. (2010), DP was only 138 manufactured by Anpon Corporation in Huai'an, China, which is 300 to 500 km far away 139 from target bays to the south, and there has been no manufactory reported near Jiaozhou, 140 Taozi and Sishili Bay.

141 Significant differences of the DP distributions have been found between Jiaozhou Bay 142 and Taozi and Sishili Bay (student *t*-test, p<0.05). In Jiaozhou Bay, the highest concentration 143 (187 pg/g dw) was found at the estuary of Haibo River, where one of Qingdao waste water 144 treatment plants (WWTP) is located nearby. The concentration at this site was one to three 145 orders of magnitudes higher than those at other sites. Guardia et al. (2010) detected DP in 146 sewage sludge from a WWTP in USA ranging from 112 to175 ng/g TOC with the content of 147 the TOC ranging from 7% to 28% (Guardia et al., 2010). Assuming all of the sewage sludge 148 contained 7% TOC, the calculated lowest DP concentration should be 7.8 to 12.3 ng/g dw. It 149 was higher than most of reported concentrations in sediments except Lake Ontario (Canada), 150 which indicates that the WWTP might be a major DP source for adjacent area. The three 151 highest concentrations of DP in the coastal sediments of Jiaozhou Bay were found near the 152 Haibo estuary, which may be influenced by the WWTP, too. In riverine sediments from the

Dague River, the DP concentration in the upstream sediment (3.0 pg/g dw) was one order of magnitude lower than that in the downstream sediment (56.9 pg/g dw). Discharge from the Jiaozhou town (36.17° N, 120.00° E) located near Dagu River might be the local source of DP.

157 Concentration of DP in the sediment collected near the Qingdao Harbour, which was the 158 third busiest habour in China, was 31.8 pg/g dw, however, in farther locations, the levels were 159 one order of magnitude lower (4.0 and 2.0 pg/g dw).

160 In Taozi and Sishili Bay, the mean concentrations of DP (33.8 and 64.5 pg/g dw, 161 respectively) were higher than that of Jiaozhou Bay (23.7 pg/g dw). In Sishili Bay, high 162 concentrations were not found along the coastal zone but in the inner bay. Ren et al. (2008) reported significant correlation between the airborne DP concentration and the population 163 164 over one million in one city. Yantai City has a population of 1 800 000 which is smaller than 165 that of Qingdao City (a population of 2 296 000). The DP concentration in Yantai should be 166 lower than that of Qingdao, which indicated the high levels in the sediments may not mainly 167 come from the local atmospheric deposition. Harbors were also suspected as possible source 168 because of intensive human activities. Yantai Harbor is settled at the west coast of Sishili 169 Bay, but it is smaller than the Qingdao Harbor (southwest of Jiaozhou Bay), which means 170 Yantai Harbor may not contribute much to the DP contamination in Sishili Bay. There was no 171 intensive industrial source reported in this area, either. The only suspected local source is the 172 WWTP settled at the Zhifu Island. This WWTP has limit ability to treat the city domestic and 173 industrial wastewater, and discharges waste water directly into the sea.

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175 3.2 Franctional abundances of Dechlorane Plus isomers

The f_{syn} value was calculated as syn-DP/(syn-DP+anti-DP) (Qi et al., 2010). The mean f_{syn} values for Jiaozhou, Taizi and Sishili Bay were 0.29 ± 0.06 , 0.18 ± 0.05 , and 0.15 ± 0.07 , respectively. Wang et al. (2010) reported the f_{syn} for commercial DP mixture to be 0.40 in 179 China which is higher than the DP ($f_{syn} = 0.2$ to 0.35) produced and widely used in North 180 America(Hoh et al., 2006; Tomy et al., 2007). In Chinese coastal sediments, the values of f_{syn} 181 were all below 0.40, and only 8 out of 48 samples showed $\geq 0.35 f_{syn}$ values. This indicates 182 definite stereo selective degradation of DP and the *anti*-DP seems more stable than *syn*-DP in 183 the coastal sediments.

In the semi-closed Jiaozhou Bay, as shown in Table 1, most of f_{syn} values (0.21 to 0.39, exclude one site with value of 0.11) were close to the technical DP mixture. They were comparable with that of Huai'an canal sediments (0.24 to 0.30) which were close to the manufacture area (Wang et al., 2010b). The DP in this bay should come from the input of the technical mixture from the adjacent region.

189 Both Sverko et al. (2007) and Tomy et al. (2008) found that the f_{syn} in the Lake Ontario 190 was below 0.20, but the concentrations of total DP were much higher than that of the 191 sediments from other places, e. g. Lake Winnepig $(30.0\pm3.2 \text{ pg/g dw})$ and Lake Erie (0.061 to 192 8.62 ng/g dw)(Tomy et al., 2007; Sverko et al., 2008). The same trend was obvious in this 193 study. In Sishili and Taozi Bay, the concentrations were significantly correlated with the f_{svn} 194 values (r = 0.58, n=19, p<0.05), and the higher concentration was accompanied by the lower f_{syn} values. As opposed to Jiaozhou Bay, Sishili and Taozi Bay are more open. Coastal 195 196 sediments from these bays may not only receive input from land, but also the inner Bohai Sea 197 (Cheng and Gao, 2000). As shown in Fig. 4, the two bays were at the south-east of the Bohai 198 strait. Marine current runs though the narrow strait with a high speed carrying the sediments 199 from the Bohai Sea. When it comes to the Yellow Sea, it slows down at the spacious area. As 200 a result, some suspended sediments settle down, and some bottom sediments stop moving. 201 The arrows in Fig. 4 show the main direction of the sediment movement which supported that 202 Taozi and Sishili Bay receive sediment from Bohai Sea. Sverko et al. (2008) reported f_{syn} 203 below 0.1 in Niagara and Lake Ontario non-depositional sediments. This suggests that the foreign sediment would enrich anti–DP and decrease the f_{syn} value. The Bohai Sea is a hot 204

spot for POPs study. High concentrations of PAHs, PBDEs and PCBs have been found in the sediments (Pan et al., 2010; Zhao et al., 2005; Liu et al., 2009). Though there is no DP data in Bohai Sea reported, high concentration of DP can be supposed as it receives large input of contaminations from surrounding cities. More works should be conducted to figure out the sources of DP in Bohai Sea.

210 *3.3 Dechlorinated species of Dechlorane Plus*

Sverko et al. (2008) reported DP decomposition in the environmental compartments. To 211 212 guarantee the ion fragment clusters truly come from the environmental samples but not the 213 artifices during analysis procedure, a new GC liner was applied before injection. Moreover, 214 one 10 pg/µL quality control standard (including *syn*-DP, *anti*-DP, aCl10DP and aCl11DP) 215 was analyzed after every eight samples to supervise the possible degradation in GC-MS 216 system. The relative deviation of all target compounds in QC standards were within 15% of 217 the original standards in all cases, which means no obvious decomposition of syn- and anti-218 DP had happened in the GC-MS system. In all of the 48 samples, 9 samples showed aCl11DP 219 concentration above the MDL, whereas, all the samples showed aCl10DP concentration 220 below the MDL. 7 out of 13 samples were detected aCl11DP in Sishili Bay, and the 221 degradation rate ((aCl10DP+aCl11DP)/(syn-DP+anti-DP+aCl10DP+aCl11DP)%) ranged 222 from 0.9% to 5.1%. Both photodegradation and aerobic microbial degradation may happen in 223 surface sediments which result in the appearance and dechlorinated species (Sverko et al. 224 2008).

- 225 *3.4 Inventory of DP in the marine sediment*
- The inventory of DP in sediment in the bays was calculated by the equation as follow:
- 227 Inventory = $C \times \rho \times A \times D \times a$

where C (pg/g) is the mean concentration of DP in the bay, ρ (g/cm³) is the dry density of

the sediment. A (cm^2) represents the area of the bay, D (cm/a) is the sedimentation rate and a

230 (a) is years which are interesting in this study.

231 The parameters for the three bays were presented in Table 3. The sedimentation rate of the three bays were close to 1 cm/a, so it is estimated for five-year inventories. They were 0.33 kg 232 233 in Jiaozhou Bay, 0.37 kg in Taozi Bay and 0.49 kg in Sishili Bay. The smallest Sishili Bay 234 (130 km²) stores most of DP (0.49 kg). Qiu et al. (2007) reported the inventory of DP per area (9cm and 16cm depth) was 120 ng/cm² in Lake Ontario, and Wang et al. (2010) calculated the 235 20.3±7.5 cm soil inventory to be 15000 to 110000 ng/cm² in Huai'an. The 5 cm depth 236 burdens of DP in per area were 8.5 ng/cm² in Jiaozhou Bay, 20.1 ng/cm² in Taozi Bay and 237 37.7 ng/cm² in Sishili Bay, respectively. Sediment would store DP for a long time as one of 238 239 the possible major sink(Wang et al., 2010). With a half-live of 14 years, DP is bioavailable for 240 many aquatic organisms, and it might be harmful especially to the benthic biota (Ismail et al., 241 2009).

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243 **4.** Conclusions

244 DP was detected in all surface sediments from three Chinese bays, but the different 245 distribution patterns and sources were found. In semi-closed Jiaozhou Bay, higher 246 concentrations were detected near the WWTP which was considered as the major DP source. 247 Usage of commercial DP products resulted in the contamination in Jiaozhou Bay. In contrast, high DP concentrations were found in the inner places of Sishili Bay with low f_{syn} . DP in 248 249 Taozi and Sishili Bay may come from the transportation from Bohai Sea and WWTP at Zhifu 250 Island could elevate DP level in Sishili Bay. The five-year inventories in the three bays 251 indicated Sishili Bay stores more DP than other two bays.

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259 **References**

- 260 Cheng, P., Gao, S., 2000. Net Sediment Transport Patterns Over the Northwestern Yellow Sea,
- 261 Based uopn Grain Size Trend Analysis. Oceanol.Limnol. Sinica. 31, in Chinese.
- 262 Gauthier, L.T., Hebert, C.E., Weseloh, D.V.C., Letcher, R.J., 2007. Current-use flame
- retardants in the eggs of herring gulls (Larus argentatus) from the Laurentian Great lakes.
- 264 Environ. Sci. Technol. 41, 4561-4567.
- 265 Guardia, M.J.L., Hale, R.C., Harvey, E., Chen, D., 2010. Flame-Retardants and Other
- 266 Organohalogens Detected in Sewage Sludge by Electron Capture Negative Ion Mass
- 267 Spectrometry. Environ. Sci. Technol. 44, 4658-4664.
- 268 Guzzella, L., Roscioli, C., Binelli, A., 2008. Contamination by polybrominated diphenyl
- ethers of sediments from the Lake Maggiore basin (Italy and Switzerland). Chemosphere 73,1684-1691.
- Hoh, E., Zhu, L., Hites, R.A., 2006. Dechlorane Plus, a Chlorinated Flame Retardant, in the
- 272 Great Lakes. Environ. Sci. Technol. 40, 1184-1189.
- Hung, C.-C., Gong, G.-C., Jiann, K.-T., Yeager, K.M., Santschi, P.H., Wade, T.L., Sericano,
- 274 J.L., Hsieh, H.-L., 2006. Relationship between carbonaceous materials and polychlorinated
- biphenyls (PCBs) in the sediments of the Danshui River and adjacent coastal areas, Taiwan.
- 276 Chemosphere 65, 1452-1461.
- 277 Ismail, N., Gewurtz, S.B., Pleskach, K., Whittle, D.M., Helm, P.A., Marvin, C.H., Tomy,
- 278 G.T., 2009. Brominated and chlorinated flame retardants in Lake Ontario, Canada, lake trout
- 279 (Salvelinus namaycush) between 1979 and 2004 and possible influences of food-web changes.
- 280 Environ. Toxicol. Chem. 28, 910-920.
- 281 Kang, J.-H., Kim, J.-C., Jin, G.-Z., Park, H., Baek, S.-Y., Chang, Y.-S., 2010. Detection of
- 282 Dechlorane Plus in fish from urban-industrial rivers. Chemosphere 79, 850-854.

- Li, F., Song, J., Li, X., Wang, Y., Qi, J., 2003. Modern Sedimentation Rate and Flux in the
- Jiaozhou Bay. Mar. Geol. Quaternary Geol. 23, 29-33, in Chinese.
- Liu, A., Lang, Y., Xue, L., Liu, J., 2009. Ecological risk analysis of polycyclic aromatic
- hydrocarbons (PAHs) in surface sediments from Laizhou Bay. Environ. Monit. Assess. 159,
 429-436.
- 288 Minh, N.H., Minh, T.B., Kajiwara, N., Kunisue, T., Iwata, H., Viet, P.H., Cam Tu, N.P.,
- 289 Tuyen, B.C., Tanabe, S., 2007. Pollution sources and occurrences of selected persistent
- 290 organic pollutants (POPs) in sediments of the Mekong River delta, South Vietnam.
- 291 Chemosphere 67, 1794-1801.
- 292 Moon, H.-B., Kannan, K., Choi, M., Choi, H.-G., 2007a. Polybrominated diphenyl ethers
- 293 (PBDEs) in marine sediments from industrialized bays of Korea. Mar. Pollut. Bull. 54, 1402-
- **294** 1412.
- 295 Moon, H.-B., Kannan, K., Lee, S.-J., Choi, M., 2007b. Polybrominated diphenyl ethers
- 296 (PBDEs) in sediment and bivalves from Korean coastal waters. Chemosphere 66, 243-251.
- 297 Palm, A., Cousins, I.T., Mackay, D., Tysklind, M., Metcalfe, C., Alaee, M., 2002. Assessing
- the environmental fate of chemicals of emerging concern: a case study of the polybrominated
- diphenyl ethers. Environ. Pollut. 117, 195-213.
- 300 Pan, X., Tang, J., Li, J., Guo, Z., Zhang, G., 2010. Levels and distributions of PBDEs and
- 301 PCBs in sediments of the Bohai Sea, North China. J. Environ. Monitor. 12, 1234-1241.
- 302 Qi, H., Liu, L., Jia, H., Li, Y.-F., Ren, N.-Q., You, H., Shi, X., Fan, L., Ding, Y., 2010.
- 303 Dechlorane Plus in Surficial Water and Sediment in a Northeastern Chinese River. Environ.
- 304 Sci. Technol. 44, 2305-2308.
- 305 Qi, J., Li, F., Song, J., Gao, S., Wang, G., Cheng, P., 2004. Sedimentation Rate and Flux of
- 306 the North Yellow Sea. Mar. Geol. Quaternary. Geol. 24, 9-14, in Chinese.
- 307 Qiu, X., Marvin, C.H., Hites, R.A., 2007. Dechlorane Plus and Other Flame Retardants in a
- 308 Sediment Core from Lake Ontario. Environ. Sci. Technol. 41, 6014-6019.

- 309 Qiu, X., Zhu, T., Hu, J., Polybrominated diphenyl ethers (PBDEs) and other flame retardants
- 310 in the atmosphere and water from Taihu Lake, East China. Chemosphere. 80, 1207-1212
- 311 Ren, G., Yu, Z., Ma, S., Li, H., Peng, P., Sheng, G., Fu, J., 2009. Determination of
- 312 Dechlorane Plus in Serum from Electronics Dismantling Workers in South China. Environ.
- 313 Sci. Technol. 43, 9453-9457.
- 314 Sverko, E., Tomy, G.T., Marvin, C.H., Zaruk, D., Reiner, E., Helm, P.A., Hill, B., McCarry,
- 315 B.E., 2008. Dechlorane Plus Levels in Sediment of the Lower Great Lakes. Environ. Sci.
- 316 Technol. 42, 361-366.
- 317 Tomy, G.T., Pleskach, K., Ismail, N., Whittle, D.M., Helm, P.A., Sverko, E., Zaruk, D.,
- 318 Marvin, C.H., 2007. Isomers of Dechlorane Plus in Lake Winnipeg and Lake Ontario Food
- 319 Webs. Environ. Sci. Technol. 41, 2249-2254.
- 320 Tomy, G.T., Thomas, C.R., Zidane, T.M., Murison, K.E., Pleskach, K., Hare, J., Arsenault, G.,
- 321 Marvin, C.H., Sverko, E., 2008. Examination of isomer specific bioaccumulation parameters
- 322 and potential in vivo hepatic metabolites of syn- and anti-Dechlorane Plus isomers in juvenile
- 323 rainbow trout (Oncorhynchus mykiss). Environ. Sci. Technol. 42, 5562-5567.
- 324 US, E., 2008. U.S. Environmetal Protection Agency. HPV Voluntary Robust Summary.
- Wang, B., Iino, F., Huang, J., Lu, Y., Yu, G., Morita, M., 2010a. Dechlorane Plus pollution
- and inventory in soil of Huai'an City, China. Chemosphere 80, 1285-1290.
- 327 Wang, D.-G., Yang, M., Qi, H., Sverko, E., Ma, W.-L., Li, Y.-F., Alaee, M., Reiner, E.J.,
- 328 Shen, L., 2010b. An Asia-Specific Source of Dechlorane Plus: Concentration, Isomer Profiles,
- and Other Related Compounds. Environ. Sci. Technol. 44, 6608-6613.
- 330 Wu, J.-P., Zhang, Y., Luo, X.-J., Wang, J., Chen, S.-J., Guan, Y.-T., Mai, B.-X., 2009.
- 331 Isomer-Specific Bioaccumulation and Trophic Transfer of Dechlorane Plus in the Freshwater
- 332 Food Web from a Highly Contaminated Site, South China. Environ. Sci. Technol. 44, 606-
- **333** 611.

334	Zhao, X., Zheng, M., Liang, L., Zhang, Q., Wang, Y., Jiang, G., 2005. Assessment of PCBs
335	and PCDD/Fs along the Chinese Bohai Sea coastline using mollusks as bioindicators. Arch.
336	Environ. Con. Tox. 49, 178-185.
 337 338 339 340 341 	Figure caption and table title
342	Figure 1. The sampling sites in Taozi, Sishili and Jiaozhou Bay
343	Figure 2. Distributions of DP in surface sediments from Jiaozhou Bay
344	Figure 3. Distributions of DP in surface sediments from Toazi and Sishili Bay
345	Figure 4. Sketch map of sediment transportation. Five-pointed stars represent some cities
346	with population over one million
347	
348	Table 1. Concentritions (pg/g dw) of syn-DP, anti-DP, $aCl_{10}DP$, and $aCl_{11}DP$ and the f_{syn}
349	value in surface sediments of Jiaozhou Bay, Taozi Bay and Sishili Bay, North China
350	
351	Table 2. Comparisons of DP level in China and other area in the world
352	
353	Table 3. Inventory parameters of Jiaozhou. Taozi and Sishili Bay. China
354 355 356 357 358 359 360 361 362 363 364 365 366 367 368 369 370 371 372 373 374	



Fig. 1 The sampling sites in Taozi, Sishili and Jiaozhou Bay



Fig. 2 Distributions of DP in surface sediments from Jiaozhou Bay



Fig. 3 Distributions of DP in surface sediments from Toazi and Sishili Bay



Fig. 4 Sketch map of sediment transportation. Five-pointed stars represent some cities with population over one million